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Assessment of the External Costs of Life Cycle of Coal: The Case Study of Southwestern China

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Abstract: Coal will continue to be the main energy source in China for the immediate future, although the environmental pollution and ecological impacts of each stage in the full life cycle of coal mining, transportation, and combustion generate large quantities of external costs. The Late Permian coals in southwestern (SW) China usually contain high amounts of fluorine (F), arsenic (As), and ash, which together with high-F clays cause abnormally high levels of endemic fluorosis, As poisoning, and lung cancer in areas where coal is mined and burned. In this paper, we estimate the external costs of the life cycle of coal. The results show that the externalities of coal in SW China are estimated at USD 73.5 billion or 284.3 USD/t, which would have accounted for 6.5 % of the provincial GDP in this area in 2018. The external cost of human health accounts for 87.2% of the total external costs, of which endemic skeletal fluorosis diseases and related lung cancers have the most important impact. Our study provides a more precise estimate of externalities compared with its counterparts in other provinces in China. Therefore, several policy recommendations would be proposed to internalize the external cost.

Keywords: coal; southwestern China; life cycle assessment; external costs; endemic diseases

1. Introduction

Coal is an important energy source, supplies about 40% of the world's electricity, and is likely to continue to provide significant amounts of energy in the foreseeable future [1]. The increasing demand for electricity in southern and southeastern Asia, including China, India, and Indonesia, will continue to cause an increase in the use of coal. Coal consumption currently accounts for 59% of all energy consumed in China with its production and consumption continuing to grow in recent years [2]. Thus, finding sources of efficient and clean energy is a primary concern when addressing the severe environmental pollution and ecological damage caused by the use of coal energy.

The mining, transportation, and combustion of coal are generally harmful to the environment and human health. Many fatal accidents have been caused by underground water leakage, poisoning, and methane explosions during underground mining. For example, over 100,000 coal miners have been killed in the United States since 1900 [3]. In China, coal mine fatalities fell from 2630 in 2009 to 333 in 2018 so that mining of one million tons of coal resulted in an average of 0.093 deaths in 2018 [2]. Inhalation of dust generated during underground mining often causes coal workers' pneumoconiosis (CWP; black lung disease). In 2018, about 19,468 cases of CWP were reported in



China [4]. Mountaintop removal (MTR) mining has negative impacts on carbon storage, water quality, air pollution, and the biosphere [5]. Direct hazards caused by the transportation of coal include road dust, death and injuries from accidents, air pollution, and greenhouse gas (GHG) emissions [5]. Coal combustion releases many harmful chemicals into the environment exposing people to potentially harmful effluents: this problem is particularly obvious in China [6]. For example, endemic fluorosis is only discovered in southwestern (SW) China [7]. Moreover, the by-products of coal combustion include CO₂, methane, PM 2.5, nitrogen oxides (NOx), and sulfur oxides (SOx), which will cause air pollution, acid precipitation, and climate change through the effects of GHGs created by burning coal [5]. Therefore, these negative impacts exert additional external costs on coal use.

External costs or externalities indicate to the unaccounted and uncompensated impacts of a production process imposed on society or environment that cannot be reflected from the market pricing [8,9]. External costs should reflect the monetized value of the damage caused by the life cycle of coal mining, transportation, and combustion on human health, the environment, streams, forests, and ecosystems [10]. These external environmental costs have received much attention from many researchers. Epstein et al. [5] reviewed the life cycle of coal and estimated the external costs of coal-fired power generation in the Appalachian region in the US. Rabl and Spadaro [11] presented the results of the external costs of the combustion of fossil fuels that caused significant air pollutants. Spath et al. [12] provided a detailed technical report with a life cycle assessment (LCA) related to the production of electricity from coal that was performed in order to examine the environmental aspects of current and future pulverized coal boiler systems. Mahapatra et al. [10] performed an LCA on the production of electricity from conventional coal-based electricity generating systems in two cities in western India. These studies attempted to estimate the costs of damages and assigned monetary values to human health, crops, and building materials by using dose-response functions. Karkour et al. [13] evaluated the recent external cost of electricity generation in G20 countries using a global LCA method. The above studies provided the externalities of coal-fired electricity generation. Jorli et al. [14] further estimated and compared the external costs from fossil electricity generation from 13 developing countries. Moreover, some authors further suggested post-tax subsidies which reflect the gap between consumer prices and economically efficient prices [15-17] and proposed to carry out the reform of the electricity environmental protection price [18] based on the externality of coal in China.

Nevertheless, external costs of coal in China have received little attention. A pioneer research published in 2003 estimated huge costs spent in maintaining health and the environment from the air pollution caused by coal-fired power plants in Shandong province [19]. Dang et al. [20] calculated the economic losses caused by environmental pollution and ecological damage from coal mining activities in Shanxi province. Costs related to the ecological damage due to coal mining in this region were also calculated by Zhang and Lian [21] and Cao [22]. Feng et al. [23] calculated the atmospheric damage caused by air pollution from coal combustion in Fuxin, Liaoning, China. In these studies, the external environmental costs of coal included costs of losses associated with environmental pollution and ecological damage. However, most research conducted in China is related to the environmental costs of coal that only occupies a certain stage of the coal chain, while the life cycle of coal is not considered enough. For example, coal mining and transportation were not covered to reflect the true external environmental costs. Moreover, ecological damage and human impacts aspects have usually been overlooked. Recently, a few studies have been published based on life cycle analysis of the coal chain in China. Liu et al. [24] estimated the external environmental cost of coal for its full life cycle in China. Zhang [25] and Lv et al. [26] calculated the external environmental costs of coal in Shanxi and Inner Mongolia, respectively. Wang et al. [27] used the LCA method to analyze how coal-fired power generation impacted the environment in China. These authors established an accounting system including costs for environmental pollution, the ecological damage, and impacts to human health during the life cycle of coal. The environmental externalities for biomass- and coal-fired electricity [9,28], flue gas cleaning processes, and carbon emissions of coal-fired power generation [29,30] in China were

also assessed. SW China is the only region that endemic fluorosis is discovered for coal burning all around world. However, there is little knowledge that has been shared about the external costs of coal in SW China.

To fill in the aforementioned gap, comprehensively calculated external costs of the life cycle of coal are needed to focus on coal mining, transportation, and combustion to indicate the true cost of coal use in SW China. In this paper, we aim to (1) monetize the external costs using the life cycle of coal in SW China for the first time; (2) compare our results with other researches in order to reveal the main sources of external costs; and (3) propose helpful policy recommendations.

2. Materials and Methods

2.1. Study Site

Coal has played an important role in China's economic development especially in SW China (Yunnan, Sichuan, and Guizhou provinces), where coal was relied heavily on as a means of economic development (Figure 1), despite concerns related to air pollution and GHG emissions. China has coal reserves of about 1.67 trillion tons [31], while SW China has reserves of about 0.11 trillion tons [32]. Coal production and consumption account for a large proportion of total energy produced in SW China, although these have been declining for the past decade (Table 1) in these provinces. SW China is blessed with abundant hydroelectric power resources accounting for about 80% of the total undeveloped capacity of hydropower in China [33], which can provide a good alternative energy source in the future.

Year	Sichuan /10 ⁴ Ton		Guizhou /10 ⁴ Ton		Yunnan /10 ⁴ Ton	
	Production	Consumption	Production	Consumption	Production	Consumption
2010	9247.93	11,520.40	15,954.02	15,601.02	5875.86	4920.86
2011	9377.47	11,454.34	11,177.47	12,084.97	6239.15	5339.53
2012	9470.70	11,872.20	18,107.05	14,048.96	6275.75	5546.57
2013	7574.90	11,658.60	18,517.78	13,650.74	6829.46	5099.27
2014	7662.80	11,045.39	18,508.29	13,117.60	2611.39	4502.65
2015	6406.46	9288.90	17,204.99	12,833.49	3319.57	4245.48
2016	6164.79	8869.49	16,850.64	13,642.75	2822.04	4287.91
2017	4798.51	7855.88	16,343.80	13,409.76	4674.98	4088.13

Table 1. Coal production and consumption in southwestern China from 2010 to 2017.

Source: Statistical Bureau of Sichuan (2018), Statistical Bureau of Guizhou (2018), and Statistical Bureau of Yunnan (2018) [34–36]. Reproduced from China Statistics Press: 2018.

The Late Permian is an important period of coal accumulation in geological history [37]. During this period, rich coal resources formed in SW China, and these became an important coal base [37]. The coal-bearing basins are located in the western part of the Yangtze Block. The Late Permian marine incursion changed the paleoenvironments from terrigenous to marine in the western Yangtze Block and formed most of the coal sequences [38], including the Changxing and Longtan formations of that period [39]. High ash yields in these coals are attributed to the high detrital input [40]. High detrital input with marine influence, volcanic ash input, and high hydrothermal solution is the primary geological factor causing trace element anomalies in coal in SW China [41–43]. Thus, geologically, the types of Permian coal in SW China typically differ from coal found in other provinces.





Figure 1. The major coal bases in China (Modified from Dai et al. [44]). 1—Shendong; 2—Northern Shaanxi; 3—Eastern Ningxia; 4—Huanglong; 5—Northern Shanxi; 6—Central Shanxi; 7—Eastern Shanxi; 8—Central Hebei; 9—Southwestern Shandong; 10—Henan; 11—Lianghuai; 12—Northeast China; 13—Yungui.

Coal contains almost all of the elements found in the periodic table. In particular, F, As, Hg, Cd, Pb, and U are usually considered to be toxic to human health and in the environment [6,43]. The F content in Chinese coals averages 130 μ g/g [43]. Endemic fluorosis has been reported in Guizhou and Yunnan provinces [6,43,45], which is caused by the burning of coal and clay with high F levels resulting in the closure of some mines [6]. Endemic fluorosis related to coal burning is only discovered in China. Clay with the F content averaging 1027.6 mg/kg is used as an additive for coal-burning or as a briquette binder for fine coal [46], which is much higher than those in western Guizhou and other Chinese coals [43]. A recent study has shown that As in coal is highly enriched in southwestern Yunnan and part of Guizhou [47]. Endemic arsenism caused by domestic As-rich coal combustion is locally very severe in southwestern Guizhou [43]. Coals with high Hg content occur in SW China including in the coalfields of Guizhou and Yunnan provinces [48]. The arithmetic mean of cadmium in Chinese coals is 0.43 μ g/g, while Cr is enriched in Sichuan province with a mean content of 5.91 μ g/g [49]. The Pb content in coals of SW China averaged $25.54-29.51 \mu g/g$ [50], higher than the mean value of Dai et al. [43]. Almost all of the U-rich coals in China come from southwestern and northwestern China [51]: the mean U content was 200–229 and 153–167 mg/kg in the Guiding coalfield in Guizhou and the Yanshan coalfield in Yunnan, respectively [41,51,52]. Beyond that, indoor emissions from household coal burning in Xuanwei County in Yunnan caused this region to have the highest female lung cancer mortality rates in China [53], which was attributed to nanometer-scale quartz in coals [54].

As a result, most types of coal found in SW China are rich in toxic trace elements including F, As, Hg, Cd, Pb, and U. Endemic diseases such as fluorosis, arsenism, and lung cancer are mainly restricted in this region of China, which not only increase the potential health risk but also generate a large number of external environmental costs. However, no known published papers were found that assessed the externalities of coals in this type of area.

2.2. Methods

Life cycle analysis that examines all stages of coal production from mining to utilization calculates the full external costs and is critical to informing the public and guiding policy formulation [28]. The external costs in this study are shown in Figure 2. Usually, two kinds of methods have been

considered to assess the external environmental costs. One is based on cost control, while the other more popular approach is based on damage loss [27]. Since the 1990s, damage losses have been used to account for external environmental costs of power generation in the US and EU [27]. The damage loss method includes market value, opportunity cost, and human capital methods.



Figure 2. The external costs of coal production in SW China.

2.2.1. System Boundaries

Usually, for the coal-fired power plant, the external costs consist of three subsystems: coal mining, transportation, and electricity generation [12]. Some authors suggested five subsystems including coal mining, processing, transportation, coal-fired power generation, and resource utilization [30]. In this study, the external environmental costs are defined as costs created when the coal industry exerts a negative impact on the environment and human health including costs related to coal mining, transportation, and utilization. Coal usage consists of the combustion of coal-fired power stations and indoor burning. The ecological damage and environmental pollution, which includes negative effects on human health, are considered in all three stages of coal extraction and use. The ecological damage can be divided into impacts on forests, farmland and crops, grasslands, as well as lakes and rivers. Environmental pollution includes air, water, and solid waste pollution. The external costs of human health are composed of casualties in coal mining, transportation, and diseases related to air pollution during the life cycle of coal production.

2.2.2. Index Systems

This paper focuses on the monetized impacts and is limited by data availability based on the actual situation in SW China in 2018. Given that, only human health, environmental pollution, and ecological degradation or damage are accounted for in the external costs. Costs in crop failure caused by wastewater pollution, wetland destruction, storage of coal and related dust pollution, and tourism loss are excluded in this study. A series of accounting items of physical quantities are tabulated in Table 2.

Life Cycle	Accounting Items	Damage Loss	Accounting Index	
Underground mining	Human casualties	accidents	costs of death compensation and medical treatment	
	Air pollution	pneumoconiosis or black lung	costs of treatment of black lung, and salary losses of the patients and family members	
	Air pollution	pneumoconiosis or black lung	costs of treatment of black lung, and salary losses of the patients and family members	
	Water pollution	environmental pollution	costs of wastewater treatment	
MTR mining	Solid waste pollution	gangues storage	costs of solid waste storage and land function loss	
mining	Aquatic ecosystems	water resources destruction and soil erosion	costs of water sources loss and soil erosion controlling	
	Industrial and residential land	land subsidence and residents' relocation	costs of residents' relocation and land restoration	
	Forest	forest destruction	costs of forest tree loss, wood consumption, and ecological service value	
	Grassland	grassland destruction	costs of grassland ecological service value	
Transportation	Air pollution	dust pollution	costs of coal dust pollution	
	Human casualties	accidents	costs of death compensation and medical treatment	
	Air pollution	endemic diseases and lung cancer	costs of treatment of endemic diseases	
Combustion	Air pollution	human welfare loss	costs of human welfare loss	
Combustion	Air pollution	farmland destruction and crops failure	costs of ecological restoration of farmland and crops losses	
	Climate change	greenhouse gas emissions	costs of loss of greenhouse gas emissions	

Table 2. The accounting indices of the external costs of the life cycle of coal.

2.2.3. Accounting Model

(a) Air pollution cost

During the coal mining stage, black lung disease or pneumoconiosis can often lead to chronic obstructive pulmonary disease and is the primary occupational disease affecting mine workers. Some endemic diseases including fluorosis, arsenism, and lung cancer are the primary diseases caused by the burning of coal in SW China. The human capital method was adopted [26,55]. In addition to these diseases, air pollution will affect human health and lead to higher morbidity and mortality in the human population of the region. Further, acid rain caused by coal combustion contaminates the soil and causes crop failure. The total cost to human health caused by air pollution was calculated using Equation (1):

$$C_{1-1} = \sum_{i=1}^{n} (P_{i} \times C_{Ai} + P_{i} \times D_{Hi} (C_{W} + C_{PA}) + P_{i} \times D_{Li} \times C_{W} + P_{Di} \times D_{i} \times C_{W}), \quad (1)$$

where C_{1-1} is the total cost to human health caused by air pollution, *i* is the *i*th type of disease, *n* is the numbers of diseases, *P*₁ is the total number of people with the *i*th type of disease, *P*_{Di} is the total number of fatalities caused by the *i*th type of disease, *C*_{Ai} is the per capita cost of treatment of the *i*th type of disease, *C*_W is the average wage per mining worker per year, *C*_{PA} is the average wage of attendants per patient, *D*_{Hi} and *D*_{Li} are the total number of days of a hospital stay and total days of incapacity to work for all patients of the *i*th type of disease, and *D*_i is the per unit indication to the losses of working years for all deceased patients of the *i*th type of disease.

The total cost for crop reduction caused by air pollution was calculated using Equation (2):

$$C_{1-2} = \sum_{i=1}^{n} (Q_i \times R_i / (1 - R_i) \times P_i),$$
(2)

where C_{1-2} is the total cost for crop failure caused by air pollution, *i* is the *i*th kind of crop, *n* is the numbers of crops, Q_i is the annual yield of the *i*th kind of crop with air pollution, R_i is the rate of yield reduction of the *i*th kind of crop (%), and P_i is the market value per kilogram of the *i*th kind of crop.

(b) Death and injury costs

For deaths and injuries caused directly by accidents during the mining and transportation of coal, the human capital method was adopted. The present study only considered work-related fatalities because limited information was available related to injuries. The cost of such deaths was calculated using Equation (3):

$$C_2 = P_{\rm D} \times C_{\rm P},\tag{3}$$

where C_2 is the total cost caused by fatalities caused by direct accidents during the mining and transportation of coal, P_D is the total number of such fatalities, and C_P is the per compensation unit that was paid to surviving relatives in every case of work-related death based on laws.

(c) Dust pollution costs

The cost of dust pollution caused by the transportation of coal trucks and movement of coal by water was considered in SW China. Coal dust caused by the transportation, loading, and unloading of coal will make the skin of people as well as the walls and furniture in homes dirty. The restoration or replacement cost approach of Lv et al. [26] was adopted to calculate the related costs using Equation (4):

$$C_3 = C_{\rm PW} \left(P_{\rm Ti} \times K_{\rm i} + P_{\rm Ti} \times K_{\rm Li} \right), \tag{4}$$

where C_3 is the cost of damage caused by coal dust pollution, *i* is the *i*th kind of transportation type, P_{Ti} is the transportation capacity, C_{PW} is the expense related to general dust emissions, K_i is the pollution coefficient of coal dust, and K_{Li} is the production coefficient of coal dust during loading and unloading.

(d) Water pollution costs

Mountaintop removal mining without careful mitigation often contaminates downstream waters and increases the costs of wastewater treatments. The restoration or replacement cost approach was adopted to analyze these costs in Equation (5):

$$C_4 = P_W \times C_W,\tag{5}$$

where C_4 is the cost of wastewater treatment, P_W is the total water consumption of the mining industry, and C_W is the cost of wastewater treatment per m³.

(e) Solid waste pollution costs

Coal gangue increases the cost of coal mining by causing negative effects on the landscape near a mine. The shadow price approach was adopted to calculate the cost of managing gangue using Equation (6):

$$C_5 = P_{\rm ST} \times C_{\rm ST},\tag{6}$$

where C_5 is the external cost of managing coal gangue, P_{ST} is the total volume of coal gangue, and C_{ST} is the shadow price per unit volume.

(f) Effects on aquatic ecosystems

Coal mining causes water pollution and soil erosion, which can increase the external expenses of coal mining. The shadow price approach was adopted for water and soil conservation and water resource value loss as calculated in Equation (7):

$$C_6 = P_A \times A_A \times C_A + P_A \times A_R \times C_R, \tag{7}$$

where C_6 is the cost of negative effects to aquatic ecosystems, P_A is the total volume of coal extracted, A_A is the volume of soil erosion caused by each ton of coal mined (m²/ton), C_A is the cost per unit of water and soil conservation, A_R is the amount of underground water polluted for each ton of coal mined (m³/ton), and C_R is the shadow price per unit of protecting water resources.

(g) Effects on forests

Coal mining operations consume large amounts of pit wood and also displace forests, which will not only reduce tree growth but also affect the value of ecological services provided by forests. The market values method was adopted to calculate this indirect cost of mining using Equation (8):

$$C_7 = P_A \times A_F \times C_F + A_{FT} \times C_{FT} + P_F \times A_{FT} \times C_{FS},$$
(8)

where C_7 is the cost of tree harvest and forest displacement, A_F is the unit consumption of wood for per ton of coal production (m³/t), A_{FT} is the total area of affected forest, C_F is the unit market price of wood (USD/m³), C_{FT} is the unit price of forest (USD/ha), P_F is the oxygen released per unit area of forest (t/ha), and C_{FS} is the market price of per m³ oxygen.

(h) Effects to grassland

Coal mining will occupy and displace grasslands and cause the loss of the value of ecological services provided by grasslands. The shadow price approach was adopted to estimate this change using Equation (9):

$$C_8 = A_{\rm FG} \times C_{\rm FG},\tag{9}$$

where C_8 is the total value of ecological services lost through the effects on grasslands, A_{FG} is the area of grasslands affected by mining, and C_{FG} is the shadow price per unit of the value of grassland ecological services (USD/ha).

(i) Losses of industrial and residential lands

Coal mining often caused industrial lands to subside and thus wasteland resources. Moreover, a coal mining plant often occupies and pollutes lands surrounding the mines and causes the local residents to relocate [26]. The restoration or replacement cost approach was adopted to measure this cost using Equation (10):

$$C_9 = P_A \times A_L \times C_{\rm ID} + P_A \times M_P \times C_{\rm RD},\tag{10}$$

where C_9 is the total cost of the loss of industrial and residential lands, A_L is the area of land subsidence caused by each ton of coal mined (ha/ton), M_P is the number of residents having to relocate per each ton of coal mined (person/ton), C_{ID} is the cost of land subsidence based on each ton of coal mined (USD/ha), and C_{RD} is the per unit cost of the relocation of each resident.

(j) GHG emissions

Coal mining and burning release GHGs (CH_4 , CO_2) and thus have the potential to contribute to global warming. The damage cost value method was adopted to calculate this cost using Equation (11):

$$C_{10} = C_{\rm G} \times (P_{\rm A} \times W_{\rm M} + P_{\rm C} \times W_{\rm C}), \tag{11}$$

where C_{10} is the total cost of the negative effects caused by GHG emissions, $P_{\rm C}$ is the total amount of coal burned in coal-fired plants and rural villages, $W_{\rm M}$ is the emission coefficient of methane for each ton of coal mining (m³/ton), $W_{\rm C}$ is the emission coefficient of CO₂ for each ton of coal burning, and $C_{\rm G}$ is the cost per unit of the negative effect of each ton of CO₂ equivalent released (USD/ton).

2.3. Data Sources

Statistics from the database in SW China, 2018 were used in this study, e.g., the total production of coal, waste discharge, pollutants, monetized values of patients' treatment, death compensation. The benefit transfer method was used in calculating some monetized values due to the unavailable data in SW China [56–58], e.g., the value of water and soil erosion, forest and grassland destruction. Usually, the benefit transfer method includes the unit value transfer and the function transfer [59]. Due to the similarities in economic and environmental conditions between Shanxi, Inner Mongolia, and SW China, authors preferred to perform the unit value transfer. Black lung disease or pneumoconiosis as well as endemic diseases such as fluorosis, arsenism, and lung cancer are the main diseases caused by coal mining and combustion in SW China. Data related to the number of patients with pneumoconiosis and/or endemic diseases, average medical expense, average time of hospitalization, and average wage of per coal miner were mainly obtained from the Fifth Analysis Resort of National Health Services Survey in China [60], China Public Health Yearbook [61], Sichuan Public Health Yearbook [62], and the Sichuan, Guizhou, and Yunnan statistical yearbooks produced by the statistical bureaus of Sichuan [34], Guizhou [35], and Yunnan [36], respectively. All the statistical reports provide the lowest reasonable estimates because the diseases of many people likely go unreported in the rural villages in SW China. Data related to the incidence of disease, total number of coal miners, and known populations of these provinces were used to present high estimates of the related costs. The value of medical expenses is from the Sichuan Public Health Yearbook [62]. Data related to the total water consumption and the areas of soil erosion, forest, and grassland degradation as well as the amounts of gangue storage related to coal mining are from the statistical year books [34–36].

Data related to transportation accidents are from the yearbooks of each province and China's Work Safety Yearbook [63]. Data related to the dust released during the transportation of coal was estimated based on the total volume of coal transported and then multiplied by the emission coefficient of transportation. Data related to the concentration of SO₂ and precipitation pH values are from the respective Ecological Environment Bulletins of Sichuan, Guizhou, and Yunnan published by the respective Department of Ecology and Environment [64–66].

3. Results

3.1. Coal Mining

3.1.1. Mortality

In 2018, in Sichuan, Guizhou, and Yunnan provinces, a total of 42, 17, and 13 coal miners were killed in underground mining accidents, respectively [67]. The compensation for mining-related fatalities in each province is 20 times the per capita disposable income of urban residents [68]. Thus, the total cost for human fatalities in 2018 was about USD 6.8 million (Table 3). This is a low estimate because the limited data did not allow us to calculate the costs of coal mine injuries.

3.1.2. Coal Miner Pneumoconiosis

In 2014, a total of 934 cases of coal miner pneumoconiosis (CMP) were reported in Guizhou [69]. In 2007, a total of 1996 cases of CMP were reported in Yunnan and with a mortality rate of about 32.7% [70]. In addition, 720 cases of CMP were reported in Sichuan in 2018 [71]. Based on Wang et al. [72], the number of patients reported to have CMP only accounts for 6.7% of the estimated number of CMP because many patients are reluctant to go to a hospital. They estimated the mean numbers of CMP patients at about 2.05 and 73.41 persons for coal production in one million tons in large and other enterprises, respectively. Thus, the median number for the infection of pneumoconiosis is 37.7 persons per 1 million tons of coal production. In addition, we thus calculated the number of fatalities based on a mortality rate at about 22% of the people who contract these diseases [73]. The estimated number is closer to a recent survey, which reported that about 9538 CMP patients died in 2015 in China [74]. Therefore, we estimated that the total cost of CMP in SW China was about USD 421.6 million in 2017 (Table 3).

3.1.3. Wastewater

Underground mining and MTR will produce a large quantity of wastewater resulting in a large cost for wastewater treatment. The total wastewater discharged in 2018 was about 507.9, 174, and 206.1 million tons in Sichuan, Guizhou, and Yunnan, respectively [34–36]. Coal production and processing discharged about 8.7% of the total industrial wastewater in 2017 in China [75]. We thus estimate that the total amount of wastewater discharged by coal production and processing facilities was about 77.26 million tons, and the total external cost for wastewater treatment was USD 44.1 million (Table 3).

3.1.4. Gangue

Coal mining activities produce a large amount of solid waste, most of which is gangue. The ratio of gangue to coal during coal extraction is about 13% using current technology. The total coal production in SW China was 258.4 million tons in 2018. The shadow price of gangue was USD 3.57/t [20]. Thus, we estimate that the total gangue was about 33.6 million tons, and the total external cost for gangue was about USD 120.0 million (Table 3).

3.1.5. Contaminated Underground Water and Soil Erosion

Underground mining and MTR would often pollute or displace underground water resources. About 2.54 m³ of underground water would become contaminated for every one ton of coal production [76]. It is estimated that a total of 656.3 million tons of underground water resources would be contaminated. Thus, the total external cost for underground water resource contamination was about USD 728.5 million in 2018. Moreover, coal mining would lead to an increasing trend of water and soil erosion. Based on related statistics from 2018, about 26.46 million hectares of soil experienced erosion in SW China [77]. We referred to Lv et al. [26] and concluded that a total of 63,308 ha of soil erosion were caused by coal mining with a total cost of USD 98.6 million (Table 3).

Item	Impact	Data Source	External Costs (2018 USD)
	Accident deaths	In 2018, there were 42, 17, and 13 coal miners killed in Sichuan, Guizhou, and Yunnan, respectively [67]. The per unit death compensation is 20 times the per capita disposable income of urban residents [68]. In 2018, the per capita disposable income of urban residents in Sichuan, Guizhou, and Yunnan was USD 4745, USD 4513, and USD 4784, respectively [34–36].	42 persons × 20 × USD 4745/person + 17 persons × 20 × USD 4513/person + 13 persons × 20 × USD 4784/person = USD 6.8×10^6
Coal mining	Pneumoconiosis	Coal miner pneumoconiosis (CMP)—37.7 persons with a median value for coal production in one million tons [72]. In 2018, the annual coal production was about 48, 163.4, and 47 million tons in Sichuan [34], Guizhou [35], and Yunnan [36], respectively. The mortality rate is about 22% for CMP [73]. About 171 days (0.47 years) is the average time of a hospital stay for each CMP patient [79]. The average time a CMP patient lost the capacity to work was 1.7 years [23]. The total working years lost for all fatalities from CMP is averaging about 13.32 years per person [79]. The per capita cost of treatment for CMP is USD 3444 [62]. The average wages per patient and attendant per year are USD 8063.7 and USD 8381.6 in Sichuan, USD 8279.4 and USD 8989.1 in Guizhou, and USD 7271.4 and USD 10,500 in Yunnan [34–36].	In 2018, CMP patients in Sichuan: $37.7 \times 48 = 1810$ persons, Death 1810 × 22% = 398 persons; in Guizhou: $37.7 \times 163.4 =$ 6160 persons, Death 6160 × 22% = 1355 persons; in Yunnan: $37.7 \times 47 = 1772$ persons, Death 1772 × 22% = 390 persons. Sichuan: 1412 persons × USD 3444/person + 1412 persons × 0.47 year × (USD 8063.7/year + USD 8381.6/year) + 1412 persons × 1.7 year × USD 8063.7/year + 398 persons × 13.32 year × USD 8063.7/year = USD 77.9 × 10 ⁶ ; Guizhou: 4805 persons × USD 3444/person + 4805 persons × 0.47 year × (USD 8279.4/year + USD 8989.1/year) + 4805 persons × 1.7 year × USD 8279.4/year + 1355 persons × 13.32 year × USD 8279.4/year = USD 272.6 × 10 ⁶ ; Yunnan: 1382 persons × USD 3444/person + 1382 persons × 0.47 year × (USD 7271.4/year + USD 10,500/year) + 1382 persons × 1.7 year × USD 7271.4/year + 390 persons × 13.32 year × USD 7271.4/year = USD 71.2 × 10 ⁶
	Wastewater	The total wastewater discharged in 2018 was about 507.9, 174, and 206.1 million tons in Sichuan, Guizhou, and Yunnan, respectively [34–36].	$(507.9 + 174 + 206.1) \times 8.7\% \times 10^6 \times 44/m^3/7 = USD$ 44.1×10^6

Table 3. Data and external costs of coal production in SW China in 2018.

Item	Impact	Data Source	External Costs (2018 USD)
		Coal production and processing discharged about 8.7% of the total industrial wastewater [75]. The cost of wastewater treatment per m ³ is about 4 yuan [78].	
	Gangue	The ratio of gangue to coal during coal extraction is about 13% using current technology. The shadow price of gangue was USD 3.57/t [20].	$(48 + 163.4 + 47) \times 10^6 \text{ t} \times 13\% \times \text{USD } 3.57/\text{t} = \text{USD } 120 \times 10^6$
	Contaminate underground water	About 2.54 m ³ of underground water will become contaminated for every one ton of coal production [76]. Diversion cost of the unit price of water is USD 1.11/m ³ [20].	$(48 + 163.4 + 47) \times 10^{6} \text{ t} \times 2.54 \text{m}^{3} \times \text{USD } 1.11/\text{m}^{3} = \text{USD}$ 728.5 × 10 ⁶
	Water and soil erosion	The area of soil erosion caused by one ton of coal production is 2.45 m ² , and the cost per unit of water and soil conservation is USD 1557.1/ha [26].	$(48 + 163.4 + 47) \times 10^{6} \text{ t} \times 2.45 \text{m}^{2}/10^{4} \times \text{USD } 1557.1/\text{ha} = \text{USD } 98.6 \times 10^{6}$
	Land subsidence	A subsidence area of 36 ha was formed for every 1 million tons of coal production; per unit cost of land restoration is about USD 8571.4/ha [25].	(48 + 163.4 + 47) × 36 ha × USD 8571.4/ha = USD 79.7 × 10 ⁶
	Residents' relocation	About 100 people will need to relocate for every 1 million tons coal produced [78]. Residents who must relocate are given USD 4857.1 per person [80].	(48 + 163.4 + 47) × 100 persons × USD 4857.1/person = USD 125.5 × 10 ⁶
	Forest	Extraction of 1 million tons of coal will cause a mined-out area that covers about 20 ha [78]; the mined-out area influence coefficient is 2.6 [21]; the forest coverage rate is about 53% [64–66]. The prices for timber and economic forests are USD	(48 + 163.4 + 47) × 20 ha × 2.6 × 53% = 7133.8 ha
	destruction	17,571.4/ha and USD 27,428.6/ha with an arithmetic mean value of USD 22,571.4/ha [81–84]. Wood consumption is about 0.02 m ³ for every ton of coal production [21]. The oxygen released per unit area of forest is 150 t/ha [21]. The per unit price of one forest tree is USD 97.1/m ³ [81] and of oxygen produced is valued at USD 142.9/t [84].	7133.8 ha × USD 22,571.4/ha + $0.02m^3/t \times (48 + 163.4 + 47)$ × 10 ⁶ t × USD 97.1/m ³ + 7133.8 ha × 150 t/ha × USD 142.9/t = USD 815.7 × 10 ⁶

Table 3. Cont.

Item	Impact	Data Source	External Costs (2018 USD)	
	Grassland destruction	The grassland coverage rate in Sichuan, Guizhou, and Yunnan is 22.63%, 0.41%, and 0.38% [64–66]. Per unit value of grassland ecosystem services is about USD 828.6/ha [85].	$(48 \times 22.63\% + 163.4 \times 0.41\% + 47 \times 0.38\%) \times 20$ ha $\times 2.6 \times USD$ 828.6/ha = USD 0.5×10^{6}	
Coal transportation	Accident deaths	About 56 and 315 people were killed by railroad and highway accidents in 2018 in SW China.	(56 + 315) persons × 20 × USD 4681/person = USD 34.7 × 10 ⁶	
	Dust pollution	Coal in SW China is largely transported by railway (75%) and highway (25%) [25]. In 2018, there were 34.3, 26.1, and 41.1 million tons of coal transported in Sichuan [34], Guizhou [35], and Yunnan [36], respectively. Loading and unloading one ton of coal will produce about 3.53 kg of dust [25]. The pollution coefficient of coal dust is 0.8% and 1% for railway and highway transportation, respectively, and per unit of external cost of dust is USD 214.3/t [26].	Railway: $75\% \times (34.3 \times 10^{6}t + 26.1 \times 10^{6}t \times 41.1 \times 10^{6}t) = 76.13 \times 10^{6}t$ Highway: $25\% \times (34.3 \times 10^{6}t + 26.1 \times 10^{6}t \times 41.1 \times 10^{6}t) = 25.37 \times 10^{6}t$ Dust pollution cost: $(76.13 \times 106t \times 0.8\% + 25.37 \times 106t \times 1\% + 3.53$ kg/t × $(76.13 + 25.37) \times 10^{6}$ t/1000) × USD 214.3/t = USD 261.7 × 10 ⁶	
Coal Combustion	Skeletal fluorosis	Nearly 1.8 million (0.17, 1.08, and 0.57 million in Sichuan, Guizhou, and Yunnan, respectively) skeletal fluorosis patients had been identified by the end of 2013 [61]. The median of skeletal fluorosis with moderate and severe illness is about 48.5% in Sichuan [86], 39.5% in Guizhou [87], and 35.6% in Yunnan [88]. The patients had an average hospital stay of 10 days with an additional 10 years of lost work [89].	Sichuan: 170,000 persons × USD 214.3/person + 170,000 persons × 0.27 year × USD 8063.7/year + 170,000 persons×48.5% × 10 years × USD 8063.7/year = USD 6.7×10^9 ; Guizhou: 1,080,000 persons × USD 214.3/person + 1,080,000 persons × 0.27 year × USD 8279.4/year + 1,080,000 persons × 39.5% × 10 years × USD 8279.4/year = USD 35.8 × 10 ⁹ ; Yunnan: 570,000 persons × USD 214.3/person + 570,000 persons × 0.27 year × USD 7271.4/year + 570,000 persons × 35.6 % × 10 years × USD 7271.4/year = USD 15.0 × 10 ⁹	
	Endemic arsenism	Endemic arsenism was only reported in Guizhou in SW China with 2848 patients identified [61]. The heart and skin diseases caused by endemic arsenism will cost about USD 164.3/person [61,90] with an average hospital stay of 12 days/person [61].	1424 persons × USD 164.3/person + 1424 persons × 0.033 year × (USD 8279.4/year + USD 8989.1/year) = USD 1.0×10^{6}	

Table 3. Cont.

Item	Impact	Data Source	External Costs (2018 USD)
	Lung cancer	The incidences of lung cancer in Xuanwei and Fuyuan for men and women were 482.78/100,000 and 387.98/100,000 [90]. About 50% of all endemic arsenism patients are vulnerable	Patients in Yunnan: (81.9 + 43.77) × 482.78 + (73.27 + 39.61) × 387.98= 104,466 persons; Patients in Guizhou: 2848 × 50% = 1424 persons Guizhou: 1424 persons × USD 3470.6/person + 1424 persons × 0.04
		to lung cancer in Guizhou. The per capita cost of treatment for lung cancer is about USD 3470.6, and per capita length of a hospital stay is about	year × (USD 8279.4/year + USD 8989.1/year) + 1424 persons × 1 year × USD 8279.4/year = USD 17.7 × 10 ⁶ ; Yunnan: 10,4375 persons × USD 3470.6/person + 10,4375 persons × 0.04
		Patients lost their work for about 1 year with working years lost per patient at 0.39 years [23].	year × (USD 7271.4/year + 0SD 10,500/year) + 10,4575 persons × 1 year × USD 7271.4/year + 91 persons × 0.39 year × USD 7271.4/ year = USD 1.2×10^9
	Pollutants	The emissions of SO_2 , NO_X , and PM_{10} as air pollutants have a cost of about 1618 USD/t, 1833 USD/t, and 1126 USD/t, respectively [25]. In 2018, the total volume of these emissions was 487,200 t, 451,000 t, and 272,000 t in Sichuan [64], 687,500 t, 359,669 t, and 196,800 t in Guizhou [65], as well as 384,400 t, 268,834 t, and 224,154 t in Yunnan [66], respectively.	1618 USD/t × (487,200 + 687,500 + 384,400)t + 1833 USD/t × (451,000 + 359,669 + 268,834)t + 1126 USD/t × (272,000 + 196,800 + 224,154)t = USD 5.3×10 ⁹
	Global warming	About 7–8 m ³ CH ₄ is emitted for every one ton of coal production [25], In SW China in 2018, there were 1.94 billion m ³ CH ₄ emitted, which is equivalent to 29.1 million t CO_2^{e} . About 2.494 t of CO_2 was emitted for every one ton of standard coal combustion in China's coal-fired power plants [91], total emissions of CO_2 were 194.9 million tons. The value of damages was estimated at about USD 30/ton of CO_2 equivalent (CO_2^{e}) [92].	USD 30/ton × (29.1 × 10 ⁶ t + 194.9 × 10 ⁶ t) = USD 6.7 × 10 ⁹
The exchange rate of RMB to US dollar is about 1:7 in 2018.			

3.1.6. Land Subsidence and Residents Relocation

Coal mining will create a number of mined-out areas that can result in land subsidence, which can disrupt or destroy the land resources on the surface. Further, land subsidence would often destroy villages and force local residents to be relocated. Usually, about 200 people relocate for every 1 million tons of coal production in the densely populated plain areas among China's population [78]. In view of the geomorphology and population density in SW China, this number goes to 100 people for every million tons of coal produced. Thus, the total cost of land resource loss caused by land subsidence and the relocation of residents is about USD 205.2 million (Table 3).

3.1.7. Forest Destruction

Coal mining that creates large amounts of mined-out areas will result in a loss of forest resources. Land subsidence will not only destroy areas of forest and reduce tree growth but also decrease the value of ecological services provided by forests. In 2018, about 5168ha of lands was totally mined-out. We used the mined-out area influence coefficient of 2.6 [21] and believe that about a total of 13,436.8 ha areas has been affected in SW China. Based on the forest coverage rate in SW China [64–66], we thus calculated the total affected area of forest was 7133.8 ha in 2018. Therefore, we estimate that the external costs of forest resource destruction were about USD 815.7 million in total (Table 3).

3.1.8. Grassland Destruction

Moreover, the mined-out areas caused by coal mining would destroy many areas of grassland and thus damage the value of the ecological services they provide. Based on the grassland coverage rate in SW China, the total area of affected grassland was 609.1 ha in 2018. Thus, the total loss of the grassland ecological service value was USD 0.5 million (Table 3).

3.2. Coal Transportation

3.2.1. Mortality

Coal transportation accounts for about 60% of all rail traffic and 10% of all highway freight transportation in China [45]. We assumed that the rail and highway accidents are in the same proportion as coal transportation by rail and highway. Thus, we calculated about 56 and 315 people were killed by railroad and highway accidents in 2018 [63] based on the coal extraction ratio in SW China [2] for a total external cost of USD 34.7 million (Table 3).

3.2.2. Dust Pollution

Coal in SW China is largely transported by railway (75%) and highway (25%) [25]. In 2018, there were 34.3, 26.1, and 41.1 million tons of coal transported in Sichuan [34], Guizhou [35], and Yunnan [36], respectively. Thus, the total costs of dust pollution were USD 261.7 million (Table 3).

3.3. Coal Combustion

3.3.1. Skeletal Fluorosis

Based on the latest data, about 12 million (1.03, 8.79, and 2.19 million in Sichuan, Guizhou, and Yunnan, respectively) dental fluorosis patients and nearly 1.8 million (0.17, 1.08, and 0.57 million in Sichuan, Guizhou, and Yunnan, respectively) skeletal fluorosis patients had been identified by the end of 2013 [61]. We only considered the skeletal fluorosis patients for external costs. All skeletal fluorosis patients with moderate and severe illness lost their ability to work based on Zang [89]. From the surveillance of local government, the median of skeletal fluorosis with moderate and severe illness is about 48.5% in Sichuan [86], 39.5% in Guizhou [87], and 35.6% in Yunnan [88]. The total costs of skeletal fluorosis patients in SW China are about USD 57.5 billion (Table 3).

3.3.2. Endemic Arsenism and Lung Cancer

Endemic arsenism was only reported in Guizhou in SW China with 2848 patients identified [60]; those patients had a high potential to experience lung cancer. The heart and skin diseases caused by endemic arsenism will cost about USD 164.3/person [61,93]. The cost for endemic arsenism is about USD 1.1 million (Table 3).

Besides endemic arsenism, crystalline silica and fine quartz particles from coal emissions in Xuanwei and Fuyuan counties also have a high potential to cause lung cancer [45,53]. Thus, we only accounted for the endemic arsenism patients in Guizhou (50% of all patient), while Xuanwei and Fuyuan counties are reported to have the highest incidence of female lung cancer and mortality [53]; then, we estimated the number of lung cancer patients caused by the burning of coal. The incidences of lung cancer in Xuanwei and Fuyuan for men and women were 482.78/100,000 and 387.98/100,000 [90]. The mortality rate from lung cancer in China is about 30.8 fatalities per 100,000 people [61]; in Xuanwei, there were 87.2 fatalities per 100,000 people [94]. The total cost of lung cancer caused by coal combustion was USD 1.2 billion in 2018 (Table 3).

3.3.3. Pollutants

The emissions of SO₂, NO_X, and PM₁₀ as air pollutants will affect human health. The burning of coal for industry and domestic use is the major contributor to regional air pollution in Guizhou and Yunnan provinces and contributes about 60% of these air pollutants emitted in Sichuan province [95]. In 2018, the total volume of SO₂, NO_X, and PM₁₀ emissions was 487,200 t, 451,000 t, and 272,000 t in Sichuan [64], 687,500 t, 359,669 t, and 196,800 t in Guizhou [65], as well as 384,400 t, 268,834 t, and 224,154 t in Yunnan [66], respectively. Therefore, the cost of human health caused by the burning of coal was about USD 5.3 billion (Table 3).

3.3.4. Global Warming

Climate impacts were monetized, and the value of damages was estimated at about USD 30/ton of CO_2 equivalent (CO_2^{e}) [92]. In China, about 7–8 m³ CH₄ is emitted for every one ton of coal production [25]. We only consider the CH₄ emissions during coal mining and estimated that the total emissions in 2018 were 1.94 billion m³, which is equivalent to 29.1 million t CO_2^{e} . About 2.494 t of CO_2 was emitted for every one ton of standard coal combustion in China's coal-fired power plants [91]. In SW China, coal combustion is used not only for power generation but also for heating and cooking in rural villages. The indoor combustion efficiency was about 50% of that from power plants [96]. Thus, the total coal combustion including personal use in SW China was 116.17 million tons, and the total emissions of CO_2 were 194.9 million tons. We estimate that the total costs of climate change from GHG emissions were USD 6.7 billion (Table 3).

3.4. Comprehensive Analysis on External Costs

By estimating the external costs for the life cycle of coal, our understanding of the true environmental effects of mining and burning coal in SW China could be benefited as well. The total externalities related to coal were USD 73.5 billion in 2018 (Table 4), which is equivalent to USD 284.3/t. The total external costs accounted for 6.5% of the GDP of these provinces in 2018. The external costs are far beyond the actual price of each ton of coal in China. Therefore, coal use in SW China is very costly in terms of the combined costs. Coal combustion takes up about 96% of the total external environmental costs (Figure 3A) and was about USD 70.7 billion in 2018 (Table 4). The cost of externalities in coal mining and transportation were USD 2.4 billion and USD 0.3 billion in 2018, respectively (Table 4). The costs of damages to the ecosystem include about 76% of the total costs of externalities during the coal mining stage and thus have the most serious impact (Figure 3B). The costs of pneumoconiosis caused by air pollution during the coal mining stage account for 17% of total costs. Dust pollution is the primary impact on the environment and accounts for 70% of the externalities caused by coal transportation

(Figure 3C). During the coal combustion stage, the most seriously affected of the externalities is skeletal fluorosis, accounting for 87%, with endemic arsenism and related lung cancer of 10% (Figure 3D).



Figure 3. (**A**) The external cost structure for the full life cycle of coal in SW China; (**B**) the externality structure of coal mining; (**C**) the externality structure of coal transportation; (**D**) the externality structure of coal combustion.

Life Cycle	Accounting Items	Accounting Results (2018 USD)	Costs of Different Stages (2018 USD)
Underground mining	Accidents	6,764,060	
	Pneumoconiosis or black lung	421,649,703	
	Wastewater Gangue	44,146,286 119,923,440	2,441,085,639
MTR mining	Underground water destruction Soil erosion	728,532,960 98 576 887	
	Land subsidence and residents Relocation	205,242,055	
	Forest destroy Grassland destroy	815,745,656 504,592	
Transportation	Accidents Dust pollution	34,733,020 261,667,800	296,400,820
Combustion	Skeletal fluorosis Endemic arsenism and lung cancer Pollutants Global warming	57,503,938,463 1,214,400,829 5,281,619,003 6,726,000,000	70,725,958,295
Total		73,463,444,754	73,463,444,754

Table 4. The external costs of the coal life cycle in SW China in 2018.

4. Discussion

4.1. Externalities Structure

4.1.1. Human Health

The external costs in the life cycle of coal can be categorized into three sections, i.e., human health, ecosystem pollution and destruction, and global warming. Human health includes mortality, as this is prone to appear during the process of underground mining and transportation, and common fatal diseases include pneumoconiosis, skeletal fluorosis, endemic arsenism, and lung cancers. It clearly shows that human health accounted for 87.2% of the total external costs in SW China in 2018 (Figure 4).



Figure 4. Proportion of the externalities structure of the life cycle of coal in SW China in 2018.

Underground mining accidents often cause fatalities in SW China because of the difficult geographic and geomorphic conditions in mines. The mining fatalities in SW China caused many deaths and injuries with methane-related accidents causing the most fatalities [97], although the numbers have dramatically decreased in the past ten years from 352 in 2012 to 72 in 2018 [67,98]. Most fatalities were caused by trucks during coal transportation, which results in greatly increased road hazards. Expenses in fatalities contributed to about USD 41.5 million of external cost in 2018 (Table 4). Nevertheless, decreasing the number of accidents during coal mining and increasing the safety of the rail transportation of coal are essential. The numbers of coal miners who were infected by pneumoconiosis as they were overly exposed to the harmful environment during coal mining and using took up 50% of the total pneumoconiosis patients in China [72]. With the external costs at about USD 421 million in 2018 (Table 4), it shows a 14.2% year-on-year decrease compared with 2017 in terms of the number of CMP patients in China, indicating that a significantly effective solution had been adopted to prevent and treat the disease, e.g., increasing mechanical mining, strengthening the supervision to mining industries.

As mentioned above, coal and clay are highly enriched with F and As in SW China, and thus endemic fluorosis and arsenism have been reported [6,43,45,46]. Moreover, the high content of crystalline silica or fine quartz particles in coal from northeastern Yunnan may cause lung cancer with high mortality rates [43,53]. In SW China, the highest externalities came from the damage and losses caused by endemic diseases resulting from the indoor burning of coal which lead to an external

cost of USD 58.7 billion in 2018 (Table 4), of which skeletal fluorosis accounted for 97.93% of total externalities caused by these diseases. These results are consistent with the trend of pollutant emissions in China from 2008 to 2015. Pollutant emissions of S, Hg, As, and F have increased dramatically in rural communities since 2008, which is in sharp contrast to the decrease in emissions from coal-fired power plants in China [99]. A recent survey showed that there is an increased trend of endemic skeletal fluorosis [7]. Therefore, reducing the incidence of endemic diseases is the best way to decrease the externalities in SW China. The high incidence of endemic diseases in SW China has several causes. First, these areas have rural villages in the mountainous areas with extremely poor economic conditions. The residents cannot afford clean energy to replace coal, and the coal they use has a high F and As content [45]. Second, the most important cause of the coal-related disease is that the residents usually eat food cooked indoors on open coal-fired stoves, which increases the high potential for disease. As such, the local government should help to build solar energy facilities to heat rural communities and connect natural gas pipelines to support city life. Moreover, the residents should change their lifestyles and abandon the use of open coal fires for heating food. Last but not least, the local government should implement policy to forbid using any coal with high F and As contents, which means any coal mines producing such coal must be shut down and abandoned.

4.1.2. Ecosystem Pollution and Destruction

The externalities of ecosystem pollution and destruction accounted for 10.66% of the total external costs of the life cycle of coal in SW China in 2018 (Figure 4). Ecosystem pollution and destruction includes wastewater and gangue pollution, underground water, soil, land, forest, and grassland destruction from coal mining, dust pollution from coal transportation, as well as air pollutants from coal combustion (Table 4).

Based on our calculation, the externalities of wastewater are much lower than the gangues. Since only a limited amount of data was available, only the external costs of wastewater discharged during coal production and washing were considered. The large quantity of wastewater produced by coal-fired power plants was not concluded. We only considered the storage of gangue during the coal mining stage and did not include the burning of coal or the effects of fly ash. A large quantity of fly ash is produced in coal-fired power plants [100]. The real cost of fly ash produced during coal combustion is about seven to ten times that of gangue [27]. Although an underestimate of about 90% to the real externalities was adopted in the present study, it will have little effect on our analysis due to the minimum value.

The ecological damage that occurs during coal mining was estimated at USD 1.85 billion of the external costs (Table 4). The external costs of impacts on water and forest destruction were USD 728.5 million and USD 815.7 million, respectively, accounting for 83.5% of the total ecosystem damage. The externalities caused by impacts on grassland were the lowest of total ecosystem damage. These results are related to the geographic and geomorphic conditions in SW China. Most coal mines in this region are in rural mountainous areas and thus have a more severe impact on water and forest resources than on grasslands.

The externalities of dust pollution from coal transportation are estimated at much lower than those of air pollutants during the coal combustion stage (Table 4). These external costs are extremely low when compared with those of endemic diseases (Table 4).

Sulfur dioxide produced during the process of coal combustion has the most significant impact on the acidification of waterbodies [78]. Generally, SW China has excellent air quality based on the region's physical geography and environmental capacity to disperse pollutants. Therefore, no effects of SO₂ emissions and acid precipitation caused a reduction in crop yield.

4.1.3. Global Warming

The use of coal is the main contributor to GHG emissions including CH_4 , that is mainly emitted in the coal mining stage, and CO_2 ; CO is also emitted during the burning of coal [5,27]. However, CO_2 is

the main contributor from coal use to global warming. The external costs of GHG emissions in SW China were estimated to be USD 6.7 billion, which contributed to 9.49% of total externalities (Figure 4). The International Energy Agency reported about 14.50 billion tons emissions from coal occurred worldwide in 2017. By 2020, the Chinese government committed to a 40%–45% reduction in CO₂ emissions per unit of GDP compared with emissions in 2005 [95]. To accomplish this goal, using clean coal and very efficient coal-burning processes will provide the main methods of lowering total carbon emissions. Carbon dioxide capture and storage is considered to be a promising technology for the abatement of GHG emissions [101]. The International Energy Agency called the third power plant constructed in the Waigaoqiao District of Shanghai the cleanest coal-fired power plant in the world: this power plant has clean and very efficient emission technology so that emissions from this plant are far lower than from natural gas-powered plants [102]. The technology includes an ultra-supercritical 1000 MW generator, flue gas desulfurization technology with zero energy consumption, and all-weather denitration [103,104]. This plant has been proven to be successful, and large-scale popularization and application of this technology should be implemented.

4.2. Limitations and Comparisons

This study is primarily based on the available data from a wide range of government organizations and literature. Data limitations suggested us to omit some environmental impacts. For example, the externalities of fly ash, PM 2.5, NOx, and Hg released into the environment and the effects of coal combustion on human health are not easily quantifiable. Moreover, several parameters that were involved in the mathematic model are not based on local research; these parameters should have a wide range to show the low and high estimates rather than providing a definite value. Thus, some LCA software, e.g., SimaPro, should be used.

Moreover, the monetized values and data that came from other regions made the research results less persuasive. For example, the shadow price of gangue disposal and underground water, the cost for the restoration of water, soil conservation and land per unit, as well as the data collected from Shanxi and Inner Mongolia which shows the external cost for dealing with dust pollution per unit (Table 3). However, these numbers only make up a small part of the total externalities and are much lower than their counterpart in human health. Thus, we believe that final results would not be heavily affected with these numbers.

Shanxi and Inner Mongolia are the largest coal-producing provinces in China. Zhang [25] analyzed the external costs of the coal-fired power plants in Shanxi and estimated external costs of about USD 4.6 billion or USD 1.89/kW h, which is comparable to Wang et al. [27] but is much lower than the estimate of Epstein et al. [5]. For Inner Mongolia in 2014, the total external environmental costs of coal were USD 12.5 billion or USD 22.77/t [26]. Liu et al. [24] estimated the total externalities at about USD 28.2/t in China in 2010. In this study, we estimated the external costs for the life cycle of coal in SW China were USD 284.3/t in 2018. Our study reveals a much higher cost compared with similar studies in other provinces in China. The main reason is that we considered endemic diseases in the total external costs for the first time, which are a typical characteristic side-effect of the use of coal in SW China.

5. Conclusions and Policy Recommendations

This paper estimated the total external costs for the life cycle of coal in SW China by performing damage cost theory. In 2018, the total externalities in SW China were estimated to be USD 73.5 billion or 284.3 USD/t, which accounts for 6.5% of the provincial GDP in this area. The external costs show that the true cost of using coal is far beyond the price of extraction and use alone, meaning that using coal is truly not very economical. The different costs of the life cycle of coal mining, transportation, and combustion were estimated to cost about USD 2.4 billion, USD 0.3 billion, and USD 70.7 billion, respectively. The external costs of human health account for 87.2% of the total external costs, of which endemic skeletal fluorosis diseases and related lung cancers impact the most, while the externalities of ecosystem pollution and destruction and global warming are estimated to account for 10.66% and

9.49% of the whole. The endemic diseases which only occur in SW China with high morbidity caused the much higher externalities.

The growth of coal use is expected to accelerate with the increasing demands for electricity and economic development in the rural areas of SW China. We suggest that the results would provide local governments a more serious assessment in terms of the true cost and impact of coal use.

To internalize the external cost, several policy recommendations are proposed. First, clean and highly efficient technology for coal combustion should be provided to the local coal-fired power plants in order to reduce GHG and air pollutant emissions. Second, by reducing the huge fanatical subsidies to the coal industry [17] and taxing the wastewater and gangue, the real cost of coal will be reflected in the price. By adjusting various types of coal taxes and fees, the government can implement increases in environmental taxes and fees that reflect the external cost of coal on the environment reasonably and thus achieve the goals of adjusting the energy consumption structure in China through a price transmission mechanism. Third, the electricity tariff should be priced in consideration of the external costs. Thus, the reform of the electricity environmental protection price should be gradually pursued [18]. Last but not least, the real situation shows that high externalities in SW China are related to endemic diseases caused mostly by indoor burning in rural areas. Thus, local governments should help communities in rural areas to build solar energy facilities for heating, encourage the use of enclosed coal-fired stoves indoors, and connect natural gas pipelines to limit the use of coal in cities. Moreover, the residents should radically change their lifestyle so they can avoid eating food cooked on open coal-fired stoves.

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